

Critical limits of soil water availability (CL-SWA) for forest trees – an approach based on plant water status

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Abstract

Due to climate change, heat waves and drought are expected to increase in frequency and intensity in Central Europe. Thus, assessments of critical constraints of water supply in forest trees are needed to develop adequate forest adaptation measures. We present a novel 'critical limit' approach to soil water availability (SWA) for the major central European forest tree species based on the physiological plant water status. In regards to the conductivity of trees' xylem, three thresholds of pre-dawn water potential (ψ_{wp}) were chosen, referring to (i) slight conductivity loss (10 %), (ii) critical conductivity loss (50 %) and (iii) complete conductivity loss (> 90 %). In times of drought, pre-dawn water potential relates to the soil water potential at the lowest soil depth in which plant's root system is able to deplete water resources; the 'effective rooting depth' (ERD). The critical limit of soil water availability (CL-SWA) represents the proportion of plant-available water within the variable effective rooting depth (ERD) that meets both the critical soil water potential at the lower limit of the ERD and the critical plant water status. The CL-SWA-approach can be implemented in water budget models like BROOK90.

Keywords: Climate change, drought, adaptation, critical limit, xylem conductivity, pre-dawn potential, soil water potential, effective rooting depth

Zusammenfassung

Kritische Grenzen der Bodenwasserverfügbarkeit (CL-SWA) für Waldbäume – ein Ansatz auf Grundlage des Pflanzenwasserstatus

Aufgrund des Klimawandels werden sich sowohl Häufigkeit als auch Intensität von Hitze- und Trockenperioden in Mitteleuropa erhöhen. Die Beurteilung einer kritischen Wasserversorgung von Waldbäumen ist daher eine wichtige Grundlageninformation zur Entwicklung von Anpassungsmaßnahmen. Wir stellen einen neuartigen Ansatz zur Bestimmung von kritischen Grenzen der Bodenwasserverfügbarkeit (SWA) für wichtige Baumarten vor, der auf dem physiologischen Pflanzen-Wasserhaushalt basiert. Unter Betrachtung der Xylemleitfähigkeit wurden drei Schwellenwerte des Dunkelwasserpotenzials (Pre-dawn-Potenzial, ψ_{wp}) gewählt, die (i) zu geringem Leitfähigkeitsverlust (10 %), (ii) zu kritischen Leitfähigkeitsverlust (50 %) und (iii) zu komplettem Leitfähigkeitsverlust (> 90 %) führen. In Trockenzeiten steht das Dunkelwasserpotenzial in Beziehung zum Bodenmatrixpotenzial in jener Bodentiefe, in der die Wurzelsysteme gerade noch Wasserressourcen nutzen können; dies entspricht der ‚effektiven Wurzeltiefe‘ (ERD). Die kritische Bodenwasserverfügbarkeit (CL-SWA) stellt den Anteil an nutzbarem Bodenwasser dar, das sowohl zu einem kritischen Bodenmatrixpotenzial als auch zu einem kritischen Dunkelwasserpotenzial führt. Die kritischen Grenzen der Bodenwasserverfügbarkeit (CL-SWA) können in Wasserhaushaltsmodelle, wie z. B. Brook90 implementiert werden.

Schlüsselworte: Klimawandel, Trockenheit, Anpassung, Kritische Grenze, Xylemleitfähigkeit, Pre-dawn-Potenzial, Bodenwasserpotenzial, Effektive Durchwurzelungstiefe

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Introduction

Climate modeling projections for the next 100 years suggest that global annual temperatures will increase between 2 °C and 3 °C, when using medium severity scenarios (SRES B1 and A1B, IPCC 2007). Heat waves and drought will increase in frequency, intensity and duration (Schär et al., 2004; Tebaldi et al., 2006; Beniston et al., 2007). Together with projected higher probabilities of extreme storm events (Christensen et al., 2002; Leckebusch and Ulbrich, 2004; Leckebusch et al., 2006), these climatic changes will expose German forest ecosystems to environmental conditions that differ from those experienced in the past (Jansen et al., 2008). The forestry sector wants to respond to climate change by developing adaptive forest management measures in order to maintain forest resistance and resilience ability (Bolte et al., in press).

The ongoing project 'Adaptation Strategies for Sustainable Forest Management under Changing Climatic Conditions – Decision Support System 'Forest and Climate' (DSS-WuK)' develops models and methods to assist stakeholders within the forestry and environment sectors to find adequate forest adaptation measures to climate change. This work aims to compile and utilize existing knowledge about climate change impacts and their effects on the economic and ecological services of forests within a user-friendly decision support system (DSS). The DSS reflects site variation at a regional scale due to wind climate, drought and climate-induced biotic agents based on regionalized projections (CLM – Climate Local Model, REMO – Regional Model) for the above-mentioned IPCC SRES A1B and B1 scenarios. High-resolution maps showing projected climate variation as well as model outputs (including abiotic and biotic risk level, site-growth development and economic indicators) for major tree species (European beech, Pedunculate/Sessile oak, Norway spruce,

Scots pine and Douglas fir) provide decision support to the individual stakeholder (Jansen et al., 2008).

This paper presents and discusses a method to adequately assess the drought risk for the major tree species in Germany based on a 'critical limit' approach (UN-ECE 2004) for soil water availability (SWA) that is derived from plant water status. We focus on the sapling growth stage, due to the high susceptibility of trees to water stress in this regeneration phase (Czajkowski et al., 2005).

Methods

'Critical limit' definition

In vegetation ecology, critical loads and limits were first used to assess effects of air pollution on vegetation and forest ecosystems within the UN-ECE "Convention on Long-Range Transboundary Air Pollution" (UN-ECE, 1979). A critical load is defined as the maximum exposure "below which significant harmful effects on specified sensitive elements of the environment do not occur" (UN-ECE, 2004). A critical limit is a chemical criterion (e. g. the threshold of base cation to aluminium ion ratio [Bc/AI] in the soil solution) that is reached when the critical load is exceeded.

We adopt this concept for assessments of drought impact on young trees. A critical limit of plant water status occurs when cavitations in trees' xylem lowers the (stem) xylem conductivity considerably; we have chosen three thresholds referring to (a) slight conductivity loss (10 %), (b) critical conductivity loss (50 %, Bréda, 2006; Brodribb and Cochard, 2009) and (c) complete conductivity loss (> 90 %). We suggest to predominantly use the 50 % threshold (b) as the critical limit, since this limit is related to the maximum recoverable drought stress in conifers (Brodribb and Cochard, 2009), which meets quite well the criterion of a "harmful effect". Relationships between xy-

Table 1:

Xylem water potential (MPa) for three thresholds of xylem conductivity loss compiled from different literature sources. A loss of 50% of xylem conductance is regarded as the critical limit for desiccation response (Brodribb and Cochard, 2009)

| Tree species | Xylem conductivity loss | | | |
|--|-------------------------|------------|--------|--|
| | 10 % | 50 % | > 90 % | Source |
| European beech (<i>Fagus sylvatica</i>) | 2.0 | 2.6 | 4.0 | Cochard et al., 1999; Hacke and Sauter, 1995; Lemonie et al., 2002a; Lemonie et al., 2002b; Lösch, 2001. |
| Pedunculate oak (<i>Quercus robur</i>) | 2.0 | 3.0 | 4.0 | Bréda et al., 1993; Bréda et al., 1995; Cochard et al., 1992; Cochard et al., 1996; Higgs and Wood, 1995; Lösch, 2001; Nardini and Pitt, 1999; Nardini and Tyree, 1999; Simonin et al., 1994; Tognetti et al., 1998; Tyree and Cochard, 1996 |
| Sessile oak (<i>Quercus petraea</i>) | 2.5 | 3.3 | 4.4 | |
| Red oak (<i>Quercus rubra</i>) | 1.5 | 2.3 | 3.5 | |
| Norway spruce (<i>Picea abies</i>) | 2.0 | 3.5 | 4.5 | Cochard, 1992; Lösch, 2001; Lu et al., 1996; Martínez-Vilalta et al., 2004; Mayr et al., 2002; Mayr and Cochard, 2003; Rosner et al. 2006; Bond and Kavanagh, 1999; Cochard, 1992; Martínez-Vilalta and Piñol, 2002; |
| Scots pine (<i>Pinus sylvestris</i>) | 2.5 | 3.2 | 5.3 | |
| Douglas fir (<i>Pseudotsuga menziesii</i>) | 2.5 | 3.6 | 5.0 | Cochard, 1992; Domec and Gartner, 2001; Kavanagh et al., 1999; Piñol and Sala, 2000; Sperry and Ikeda, 1997; Stout and Sala, 2003. |

lem conductivity and xylem water potential, derived for example from 'vulnerability curves' (Sperry et al., 1988; Tyree and Sperry, 1989; Kolb et al., 1996; Cochard et al., 2002; Cruiziat et al., 2002; Cochard et al., 2005), can be used for referring thresholds of conductivity loss to plant water status (Table 1). In this way, drought effects can be quantified using physiological criteria.

Critical limits of soil water availability

The pre-dawn water potential (ψ_{wp}), measured just before sunrise, determines the plant water status in equilibrium with the soil water potential (ψ_m) within the plant rooting zone (Ehlers, 1996). According to this concept, plant water status relates to the soil water potential at the lowest soil depth in which plant's root system is able to deplete water resources, e. g. 140 cm depth (Bréda et al., 1995). This rooting depth meets the definitions of the 'effective rooting depth' (ERD) according to AK Standortskartierung (1996).

The critical limit of soil water availability (CL-SWA) describes the proportion of plant-available water within the effective rooting depth that meets both the critical soil water potential at the lower limit of the ERD and the critical plant water status. For the assessment of CL-SWA, specifications of effective rooting depth are very important. The architecture of root systems is mainly influenced by the parent material, the soil type, bulk density, the chemical soil conditions, the depth of ground water and species and age of trees. Due to this complexity, the root depth and the distribution in the soil profile are critical model parameters. For the estimation of the effective rooting depth we used the linking rule from Müller (2004) and Raissi et al., (2009). This rule was modified for shallow-rooting spruce on soils with unconsolidated rock by decreasing its root-

ing depth from 80 % of those of deeper-rooting species like Scots pine and oak to only 70 % (Table 2). This is in line with numerous publications (Rastin and Urlich, 1990; Heinze et al., 2001; Lehnhardt and Brechtel, 1980; Kreutzer, 1961; Bolkenius, 2001; Osenberg, 1998). For water-saturated sites (with G, A, M layers, AK Standortskartierung, 1996) we kept to the rule that the depth of the groundwater table restricts the rooting depth.

Results and discussion

Root depth estimates

Use of the root depth estimation model is illustrated using input data for precipitation from the very high resolution interpolated climate data for Germany (Hijmans et al., 2005). For the spatially distributed simulation we used the digital soil map of Germany at a scale of 1:1 000 000 (Richter et al., 2007) and the digital metadata corresponding to the above soil map. Figure 1 shows the deeper rooting activity of Scotch pine in contrast to spruce. For both trees, the rooting depth in the mountainous regions is in general 100 to 120 cm. On sandy soils derived from unconsolidated rocks, where the soil depth is much greater than in the mountain regions, under poor precipitation conditions the effective root depth reaches a depth of about 2 m, this also being the lower limit of our soil profiles.

CL-SWA is modeled using water budget models (BROOK 90 model, Federer et al., 2003) using the variables (1) tree species, (2) effective rooting depth, and (3) the hydraulic properties of the soil (e. g. Clapp and Hornberger, 1978). The parameters of the water retention curve were deduced from soil texture (Clapp and Hornberger, 1978). More details on the CL-SWA modeling will be presented in a forthcoming paper.

Table 2:

Effective rooting depth on forest sites (ERD [dm]) according to Müller (2004, modified)

| Soil substrate | Age (a) Yearly prec. (mm) | < 15 | | | | 15 - 45 | | | | 45 - 80 (100) | | | |
|----------------|------------------------------|-------|------|------|-------|---------|------|------|-------|---------------|------|------|------|
| | | > 750 | 725 | 700 | < 625 | > 750 | 725 | 700 | < 625 | > 750 | 725 | 700 | |
| Sand | Others | 7.0 | 8.0 | 9.0 | 10.0 | 11.0 | 12.0 | 13.0 | 14.0 | 15.0 | 16.0 | 18.0 | 20.0 |
| | Norway spruce | 4.9 | 5.6 | 6.3 | 7.0 | 7.7 | 8.4 | 9.1 | 9.8 | 10.5 | 11.2 | 12.6 | 14.0 |
| Loam | Others | 8.0 | 9.0 | 10.0 | 11.0 | 12.0 | 13.0 | 14.0 | 15.0 | 16.0 | 17.0 | 19.0 | 21.0 |
| | Norway spruce | 5.6 | 6.3 | 7.0 | 7.7 | 8.4 | 9.1 | 9.8 | 10.5 | 11.2 | 11.9 | 13.3 | 14.7 |
| Silt | Others | 12.0 | 13.0 | 14.0 | 15.0 | 16.0 | 17.0 | 18.0 | 19.0 | 20.0 | 21.0 | 23.0 | 25.0 |
| | Norway spruce | 8.4 | 9.1 | 9.8 | 10.5 | 11.2 | 11.9 | 12.6 | 13.3 | 14.0 | 14.7 | 16.1 | 17.5 |
| Clay | Others | 8.0 | 8.5 | 9.0 | 9.5 | 10.0 | 10.5 | 11.0 | 11.5 | 12.0 | 13.0 | 14.0 | 15.0 |
| | Norway spruce | 5.6 | 6.0 | 6.3 | 6.7 | 7.0 | 7.4 | 7.7 | 8.1 | 8.4 | 9.1 | 9.8 | 10.5 |
| Bedrock | Others | | 7.0 | | | | 9.5 | | | | 12.0 | | |
| | Norway spruce | | 6.0 | | | | 8.0 | | | | 10.0 | | |

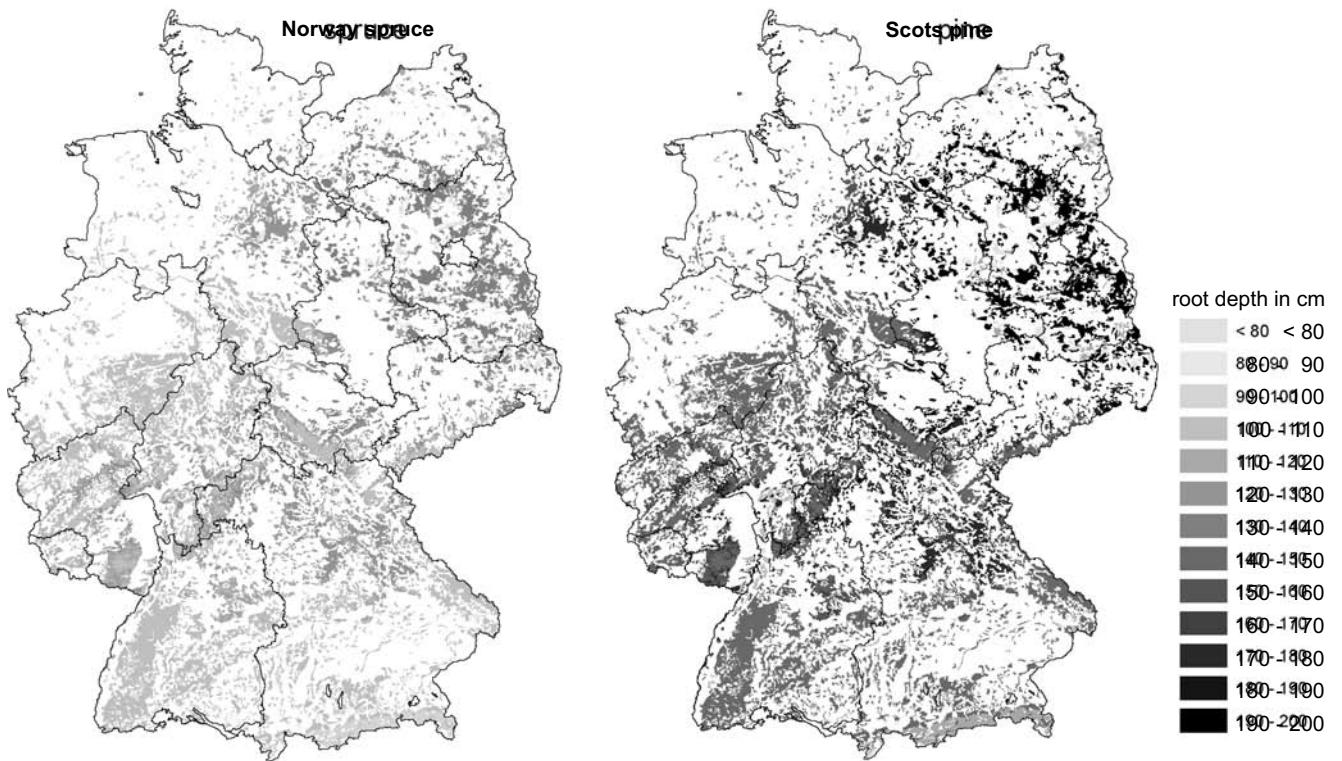


Figure 1:
Estimated effective rooting depth as a function of precipitation, soil texture and tree species

Options and constraints of the CL-SWA approach

To our knowledge, the present approach is among the first to combine the critical limit concept with plant water status and the related parameter of soil water availability. In our approach, plant water physiology is directly included in the indicator 'critical limit of soil water availability' (CL-SWA). Evident soil-plant interactions are reflected in CL-SWA indicator which determines the soil water potential that triggers a 50 % loss of xylem conductance and 95 % loss of leaf conductance. The latter indicates the maximum recoverable drought stress (Brodrribb and Cochard, 2009).

Since plant water potential is commonly used for plant drought stress in terms of xylem conductivity loss (e. g. Kolb et al., 1996; Cochard et al., 2002; Cruiziat et al., 2002; Cochard et al., 2005), the inclusion of the CL-SWA indicator in large scale water budget modelling puts physiologically based drought stress estimates on a higher spatial scale and offers novel options for forest landscape assessments.

In contrast, Novák and Havrla (2006) based their ostensibly similar estimates of critical soil water content on the decrease of water extraction ability and yield of plants; criteria which differ considerably to our criteria for CL-SWA. Moreover, they used the 'relative transpiration index' (RTI) as an indicator of the (critical) state of soil water resource

availability. RTI describes the ratio of actual transpiration to potential transpiration ($RTI = ET_a/ET_p$, Budagovskij and Grigorieva, 1991), and it is a commonly used indicator (e. g. Klap et al., 2002; Eccel et al., 2005), which may describe the stomatal control and thus adaptive behaviour of plants to water shortage rather than direct drought impact on plant water status.

There are also, however, constraints with our CL-SWA approach. One constraint is the use of pre-dawn water potential for characterizing the (critical) plant water status in order to interlink plant water and soil water status (Ehlers, 1996). Midday plant water potential, measured in times of highest evaporation pressure, often falls far below pre-dawn water potential. A correlation between the two traits is not always visible due to individual adaptation abilities to drought (e. g. Czajkowski et al. 2006). Thus, the probability of critical midday embolism is only in general reflected in the approach presented.

Another constraint is the relationship between plant water status and soil water status at the 'effective rooting depth' (ERD). The depletion depth of the plant rooting system varies due to the water-potential gradient between soil matrix potential and plant water potential (Hetsch and Heilig, 1981; Aussennac et al., 1984; Bréda et al., 1995; Tognetti et al., 1995; Lai et al., 2001; Porporato et al.,

2001; Sperry et al., 2002; Sperry and Hacke, 2002; Bhaskar, 2006). In times of extreme drought, plants seem to be able to lower the depletion depth temporarily below the ERD (Joslin and Wolfe, 2003), but this cannot be included in our approach.

The main uncertainty of our approach relates to the determination of the 'effective rooting depth' (EDR) itself, since this is highly dependent on tree species, stand structure and soil characteristics (Bibelriether, 1966; Köstner et al., 1968; Lehnhardt and Brechtel, 1980; Schmid and Kazda, 2002; Dannowski and Wurbs, 2003; Claus and George, 2005; Meier and Leuschner, 2008; Raissi et al., 2009). We lack data covering the whole variety of combinations of tree species, competition situations and soil characteristics. Thus, any general rule on ERD, like we have used for this approach and the following modeling, may consider average ERD, but not all specific situations.

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